CORONAVIRUS

Abrupt decline in tropospheric nitrogen dioxide over China after the outbreak of COVID-19

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China’s policy interventions to reduce the spread of the coronavirus disease 2019 have environmental and economic impacts. Tropospheric nitrogen dioxide indicates economic activities, as nitrogen dioxide is primarily emitted from fossil fuel consumption. Satellite measurements show a 48% drop in tropospheric nitrogen dioxide vertical column densities from the 20 days averaged before the 2020 Lunar New Year to the 20 days averaged after. This decline is 21 ± 5% larger than that from 2015 to 2019. We relate this reduction to two of the government’s actions: the announcement of the first report in each province and the date of a province’s lockdown. Both actions are associated with nearly the same magnitude of reductions. Our analysis offers insights into the unintended environmental and economic consequences through reduced economic activities.

INTRODUCTION

In December 2019, a respiratory disease, coronavirus disease 2019 (COVID-19), emerged in Wuhan City, Hubei Province, China (1). COVID-19 has since spread worldwide causing tens of thousands of deaths (2). To combat the spread of COVID-19, the Chinese government sealed off several cities reporting large numbers of infected people, including Wuhan, starting 23 January 2020; this included halting public transportation and closing local businesses. These prevention efforts quickly expanded nationwide. The policy announcements and restrictions, applied at an unprecedented scale, have implications for the Chinese environment and the economy that we quantitatively evaluate in this paper. In particular, we use satellite nitrogen dioxide (NO2) measurements to monitor changes in fossil fuel usage, related to economic activity, over China following the outbreak of COVID-19. Nitrogen oxides (NO + NO2 = NOx), emitted during high temperature combustion, are relatively short-lived in the atmosphere (lifetimes of the order of hours near the surface) and therefore remain relatively close to their sources (3). NO2 tropospheric vertical column density (TVCD) retrieved from backscattered solar radiation, such as from the Ozone Monitoring Instrument (OMI) (4), has been widely used to monitor both long- and short-term changes in fuel consumption (5, 6). OMI’s successor, the Tropospheric Monitoring Instrument (TROPOMI) (7), offers a higher spatial resolution measurement of NO2 TVCD.

RESULTS AND DISCUSSION

We observe substantial reductions of NO2 TVCD after the 2020 Lunar New Year (LNY) on 25 January 2020. Figure 1 shows 20-day averages of OMI NO2 TVCD before, during, and after the 2020 LNY (hereafter referred to as the “pre,” “peri,” and “post” periods). An average reduction of 48% in NO2 TVCD over China is observed from pre to peri periods. Consistency in the trends of retrieved NO2 TVCD is found between OMI and its successor TROPOMI (fig. S1). A reduction in NO2 TVCD is typically observed during LNY because most Chinese factories shut down for the holiday and the traffic volumes decrease, resulting in a decrease in fuel consumption and thus NOx emissions. OMI NO2 TVCD shows a pre to peri decline of 27 ± 5% (mean ± SD) from data covering the 2015 to 2019 period (fig. S2). Similarly, TROPOMI shows a reduction of 33% in 2019 (fig. S3). This suggests that the observed reduction in 2020 far exceeds (21 ± 5%) the typical holiday-related pre to peri period reduction.

Consistent with the 2015–2019 data, the 2020 NO2 TVCD 7-day moving averages show a substantial reduction during the approximately 2 weeks leading up to LNY and reach a minimum around LNY, consistent with the gradual shutdown of factories before the holiday (Fig. 2). In prior years, a rebound of NO2 TVCD usually begins around 7 days after LNY, marking the end of the holiday season. OMI and TROPOMI (fig. S4) NO2 TVCDs show similar temporal patterns before 2020 with a clear reduction before LNY and an increase shortly thereafter. However, while the 2020 data show similar initial declines in the week leading up to LNY, we do not observe the typical uptick in NO2 TVCDs starting the week after the LNY as in previous years (Fig. 2). OMI (and TROPOMI) NO2 TVCDs show a longer period of low values near the minimum. Note that the 2020 data are generally lower than previous years, probably reflecting, in part, the effects of China’s clean air policies that require installation of denitrification devices for all coal-fired power plants and cement plants (8).

To rule out the possibility that the large NO2 TVCD decreases observed in 2020 may be driven by changes in the meteorological conditions affecting local NOx chemistry and NOx transport, we use Goddard Earth Observing System coupled to the NASA Global Modeling Initiative (GEOS-GMI) (9) model simulations with constant emissions. We find that the simulated effects of meteorology on NO2 TVCD are small compared with the prolonged NO2 reduction we observe from the pre to peri periods (fig. S5). The simulation with constant emissions shows many areas with increases from the pre to peri periods (fig. S6). This suggests that in many areas, the actual
The observed NO$_2$ TVCDs.

We account for the annually varying dates of the LNY.

Moving average. Values are normalized to the mean of the pre period. Note that shading shows SE of the mean. Points are plotted at the midpoint of the 7-day provinces into account (see Eq. 1): Following the report of the first case in each province, OMI NO$_2$ TVCD declined by about 26% (coeff = −1.282, P = 0.002, Table 1, column 1).

The second policy intervention is more invasive: The government took decisive action to further reduce the spread of the virus by limiting the mobility of citizens and locking down entire provinces; on average, lockdowns occurred 3.6 days after the report of the first case. We would expect that a lockdown would be followed by a reduction in travel and business activity, which, in turn, should lead to reductions of NO$_2$ TVCD. Our model (Eq. 2) shows that OMI NO$_2$ TVCD reduces by 24% following the lockdowns (coeff = −1.073, P < 0.001, Table 1, column 2).

Last, we consider the two policies jointly (Eq. 3). We find that both the announcement of the first case reported and the lockdown are associated with a reduction in NO$_2$ TVCDs in each province (Table 1, column 3). These results suggest that the effect of the announcement is about as large (16%; coeff = −0.770, P = 0.049) as the effect of the lockdown (15%; coeff = −0.722, P < 0.001). All results are qualitatively similar using TROPOMI (table S2).

NO$_2$ reductions are closely related to improvements in air quality (10). Under normal circumstances, many Chinese cities have poor air quality that reduces life quality and expectancy (11). During the COVID-19 crisis, NO$_2$ pollution was additionally reduced by ~20% for a period of between 30 and 50 days. While temporary, these substantial reductions in air pollution may have positive health impact for lives in otherwise heavily polluted areas (12). This unusual period offers a rare counterfactual of a potential society that uses substantially less fossil fuels and has lower mobility (13).

While this research provides an early insight into the NO$_2$ changes in China in early 2020, our findings are not without limitations. Because the relationship between NO$_2$ TVCD and NO$_2$ emissions is not strictly linear, the analysis of NO$_2$ TVCD provides a qualitative description of changes in NO$_2$ emissions. Accurately quantifying the changes in NO$_2$ emissions (14) is beyond the scope of this initial assessment.

Our results suggest that the announcement of the first case was followed by a reduction in NO$_2$ emissions, with a further reduction following the actual lockdown. However, note that these results do not suggest that the mobility restrictions did not have a critical impact. Recently published work suggests that the travel restrictions in China reduced the spread of the disease by up to 80% by mid-February, in particular internationally (15). In line with our results is the finding that human mobility was reduced early on during the outbreak (16).
and may, in part, have started as early as the first case announcements, with additional reductions through lockdowns.

**MATERIALS AND METHODS**

**Satellite NO\textsubscript{2} observations**

We use retrieved NO\textsubscript{2} TVCD from both OMI and TROPOMI. OMI is a Dutch–Finnish ultraviolet–visible (UV–VIS) spectrometer (4) on board the US NASA Aura satellite that was launched in 2004. TROPOMI is a UV–VIS–near infrared (IR)–short wave IR instrument (7) on board the European Copernicus Sentinel–5 Precursor satellite that was launched in 2017. Both instruments similarly measure Earth radiance and solar irradiance spectra with spectral resolutions of approximately 0.5 nm. The ratio of radiance to irradiance at wavelengths between 400 and 496 nm is used to retrieve NO\textsubscript{2} TVCD. The ground footprint sizes are 13 km by 24 km and 3.5 km by 5.5 km (3.5 km by 7 km before August 2019) at nadir for OMI and TROPOMI, respectively. Both instruments provide nearly daily to bidaily global coverage with a local equator crossing times close to 13:30. We use the version 4.0 NASA OMI standard NO\textsubscript{2} products (17). We use the version 1.0.0 TROPOMI Level 2 offline NO\textsubscript{2} data products for 2019 and the version 1.1.0 data for 2020 (18). OMI and TROPOMI measurements are aggregated to resolutions of 0.25° × 0.25° and 0.05° × 0.05°, respectively. A given grid box value is computed by averaging the pixel-level satellite observations weighted by the amount of the pixel footprint that overlaps the grid box. We remove OMI observations with effective cloud fractions of >30% to reduce retrieval errors and those affected by the so-called “row anomaly” (19). For TROPOMI, we use only observations with quality assurance values > 0.75. For the maps shown, we calculate 20-day means of NO\textsubscript{2} TVCD around the LNY using OMI during 2015 to 2020 and TROPOMI for 2019 and 2020. We only include regions dominated by anthropogenic NO\textsubscript{x} emissions in the analysis; these are defined as regions with average annual OMI NO\textsubscript{2} TVCDs > 1 × 10\textsuperscript{15} molecules/cm\textsuperscript{2} over the period of 2005 to 2019 (fig S7) (20). For time series analysis, we further compute 7-day running averages to smooth out daily fluctuations in NO\textsubscript{2} TVCD due to retrieval noise, including the effects of clouds and influences of meteorology (wind-driven transport influences NO\textsubscript{2} TVCDs).

**Sector information**

We select facilities with reported NO\textsubscript{x} emissions > 5 Gg/year (21). The locations of 245 heavy industry plants including steel, iron, coke, oil, cement, and glass industry, and 103 power plants considered in this analysis are shown in fig S7. We compute 7-day running averages of OMI NO\textsubscript{2} TVCD for grid boxes where large power plants and other industrial plants are located for 2020 (TVCD\textsubscript{2020}) and the mean of 2015 to 2019 (TVCD\textsubscript{2015–2019}). We calculate the relative difference as (TVCD\textsubscript{2020} − TVCD\textsubscript{2015–2019}) / TVCD\textsubscript{2015–2019}.

**Table 1. Effects of the government policies on NO\textsubscript{2} TVCD.** NO\textsubscript{2} TVCD is based on OMI. We use a fixed-effects model (Eqs. 1 to 3) with the first case announced and lockdown coded as binary indicator variables. We control for the average 2015–2019 OMI NO\textsubscript{2} TVCDs to adjust for seasonal variation and include provinces’ fixed-effects to adjust for geographical variation. The “Constant” term is the average province fixed-effect used as a baseline to compare the relative effect of the policy interventions. All SEs (shown in parentheses) are clustered at the province level.

<table>
<thead>
<tr>
<th>Outcome variable</th>
<th>NO\textsubscript{2} TVCD (10\textsuperscript{15} molecules/cm\textsuperscript{2})</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>(1)</td>
</tr>
<tr>
<td>First case</td>
<td>−1.282**</td>
</tr>
<tr>
<td>announced in</td>
<td>(0.384)</td>
</tr>
<tr>
<td>province, $\beta$</td>
<td></td>
</tr>
<tr>
<td>Lockdown of</td>
<td>−1.073***</td>
</tr>
<tr>
<td>province, $\lambda$</td>
<td></td>
</tr>
<tr>
<td></td>
<td>(0.215)</td>
</tr>
<tr>
<td>Average NO\textsubscript{2}</td>
<td>0.0001</td>
</tr>
<tr>
<td>TVCD 2015–2019, $\delta$</td>
<td>(0.019)</td>
</tr>
<tr>
<td>Constant, $\alpha$</td>
<td>4.847</td>
</tr>
<tr>
<td>Number of</td>
<td>1023</td>
</tr>
<tr>
<td>observations</td>
<td></td>
</tr>
<tr>
<td>$R^2$</td>
<td>0.553</td>
</tr>
<tr>
<td>Adjusted $R^2$</td>
<td>0.539</td>
</tr>
</tbody>
</table>

*P < 0.05. **P < 0.01. ***P < 0.001.

Fig. 3. Reductions in NO\textsubscript{2} TVCDs along highways. Average TROPOMI NO\textsubscript{2} TVCD over Changchun, China (black dots) for 20 days (A) before and (B) after the 2020 LNY and (C) their difference. The locations of large power plants and other industrial plants are indicated by triangles and crosses, respectively. The lines show China National Highways.
GEOS-GMI NO2 simulations

We ran the GEOS-GMI (9) with anthropogenic and biomass burning emissions of NOx and other trace gas emissions held constant to simulate NO2 TVCD over China to estimate the potential impact of meteorology on NO2 TVCDs from January to February 2020. The simulation uses the GMI chemistry mechanism (22) and the Goddard Chemistry Aerosol Radiation and Transport component of GEOS-5 (23, 24) to interact with the GMI chemistry. The simulation’s meteorology is constrained by the Modern-Era Retrospective Analysis for Research and Applications, version 2 (25), assimilated meteorological data from the NASA Global Modeling and Assimilation Office GEOS-5 data assimilation system. The constant anthropogenic emissions are from the Representative Concentration Pathways 6.0 scenario (26) for January 2019, downscaled to higher resolution using the Emissions Database for Global Atmospheric Research version 4.3.2 (27) inventory. Constant biomass burning emissions are the January 2020 monthly mean from the Quick Fire Emissions Dataset version 2 (28). This simulation includes 72 vertical levels at a spatial resolution of 0.25° (latitude and longitude) and a model time step of 7.5 min. We sample the model output only when and where there are valid satellite observations.

Statistical analysis of policy responses

For the policy evaluation, we make use of the timing of when the Chinese government first publicly reported that a person was infected with COVID-19, which occurred on several different dates across the country’s provinces. The first public announcement of “viral pneumonia of unknown cause” in Wuhan occurred on 3 January 2020. Daily public health statements began on 11 January 2020, which included the new cases, deaths, and recoveries reported separately for each province. Of particular interest for our analysis is the time when the government announced the first case in each province (table S1). We also use the exact timing when the government put restrictive mobility policies in place, to reduce the likelihood of transmission. The first such policy was put in place for Wuhan on 23 January 2020, followed by more restrictions for other provinces shortly after (table S1).

We conduct a statistical evaluation of the exact timing of the restriction in NO2 TVCDs. While the 2020 LNY coincided roughly with the lockdown of most Chinese provinces, the government’s policy actions actually took two forms and varied over time. The first policy action was public announcements of new cases in each province, while the second policy action was to restrict movement and order citizens to stay in doors (which became known as “lockdown”). We explore the timing of these two potential candidates—announcements of new cases and restrictive mobility policies—to identify to what extent they are responsible for NO2 TVCD reductions. We take advantage of the temporal variation of these measures across the country.

To analyze the effects of these policies, we use fixed-effects models that predict tropospheric NO2 TVCD, controlling for previous years’ NO2 TVCD and fixed effects for each province

$$z_{tp} = \alpha + \beta x_{tp} + \delta z_{\text{prior}} + \nu_p + \epsilon_{tp}$$  

where $z$ is the outcome variable (daily NO2 TVCD for the period from 4 weeks before LNY to 8 weeks after LNY), $x$ is an indicator variable on and after the first case is announced on day $t$ in province $p$ (which remains 1 after the first case; otherwise coded as 0), $z_{\text{prior}}$ is the NO2 TVCD in prior years (which is the average of years 2015 and 2019 for the OMI data and of the year 2019 for the TROPOMI data where prior data are only available for 2019), $\alpha$ is the average fixed effect across all provinces, $\nu$ is the fixed effect of province $p$ (relative to $\alpha$), and $\varepsilon$ is an error term that is clustered at the province $p$.

To estimate the effect of the lockdown policy, we use the following fixed-effects model:

$$z_{tp} = \alpha + \lambda y_{tp} + \delta z_{\text{prior}} + \nu_p + \epsilon_{tp}$$

where $y$ is an indicator variable for the lockdown of the province $p$ starting on day $t$ (which is 1 during the time of the lockdown; otherwise coded as 0), and all other variables are as defined above.

We use a similar fixed-effects model predicting the effect of both policies jointly

$$z_{tp} = \alpha + \beta x_{tp} + \lambda y_{tp} + \delta z_{\text{prior}} + \nu_p + \epsilon_{tp}$$

where all variables are as previously specified. $\beta$, $\lambda$, and $\delta$ are the derived coefficients of the model.

Using the above specified fixed-effect models enables us to estimate the effect of the policy precisely, as we hold constant province-specific variation and prior year variation in NO2. Our primary analysis uses OMI data (Table 1), but our results are qualitatively unchanged if we use TROPOMI data (table S2).

SUPPLEMENTARY MATERIALS

Supplementary material for this article is available at http://advances.sciencemag.org/cgi/content/full/6/28/eabc2992/DC1

REFERENCES AND NOTES
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